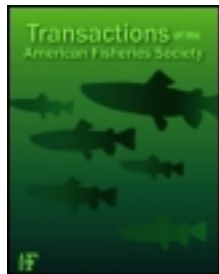


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Population Structure of a Neotropical Migratory Fish: Contrasting Perspectives from Genetics and Otolith Microchemistry

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ARTICLE

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Abstract

Developing conservation strategies for migratory fishes requires an understanding of connectivity among populations. Neotropical rivers contain diverse and economically important assemblages of migratory fishes, but little is known about the population biology of most species. We examined the population structure of *Prochilodus mariae*, an abundant migratory fish species found in Venezuelan rivers that plays essential roles in both regional fisheries and ecosystem dynamics. By coupling otolith microchemistry and microsatellite genetic analyses, we were able to evaluate both natal origins of individual fish and genetic structure on a regional level. The chemistry of otolith cores inferred separate breeding grounds for four of six populations, with 75–85% of individuals from each river sharing

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a natal signature that is distinct from the other populations. In contrast, we detected no genetic structure, indicating that gene flow among these rivers prevents population differentiation. These disparate inferences underscore the complexity of conserving migratory species; otolith data suggest that ensuring fishery sustainability requires recognizing distinct breeding stocks, while gene flow reflects the importance of connectivity across the broader river network on an evolutionary time scale. We conclude that multiple methodological approaches may often be necessary to fully understand the spatial ecology and management needs of migratory fishes and, therefore, also influence local management practices.

Understanding the population structure of migratory fishes can provide important information for determining the appropriate spatial scale for their management and conservation (Shaklee and Bentzen 1998). In ecosystems with spatially structured populations and high levels of site fidelity, heavy exploitation or habitat degradation in one location can sometimes have strong effects on that local population without affecting adjacent populations (Heath et al. 2008). Alternatively, if fish migrate widely and interbreed on a regional scale, effects on a single population could be manifested in other areas (Kell et al. 2009). Complexity, including differences among populations and their life history characteristics, can be key to maintaining ecosystem function and services (Hilborn et al. 2003; Schindler et al. 2010).

Migratory fishes display high interspecific variation in their population biology, so it is important to identify whether management and conservation efforts should treat the panoply of local populations as discrete units or instead as a single regional population. Previous studies spanning a wide range of North American taxa have identified many species with high levels of site fidelity and genetic population structure, including such diverse groups as salmonids (McDowall 2001), Swordfish *Xiphias gladius* (Alvarado Bremer et al. 2005), and Atlantic Herring *Clupea harengus* (Ruzzante et al. 2006). Other species disperse and interbreed widely and have little population structure, including Striped Bass *Morone saxatilis* (Mather et al. 2010) and Sea Lamprey *Petromyzon marinus* (Waldman et al. 2008). Population structure can also vary within species; for example, Palkovacs et al. (2008) found high levels of genetic structure for landlocked populations of Alewife *Alosa pseudoharengus*, but low levels in anadromous populations.

Using multiple methods of analyzing movement and population structure is often essential for migratory animals (Rubenstein and Hobson 2004). For fishes, chemical differences between different layers of their ear stones (otoliths) can be used to reconstruct the migration history of individual fish because otolith microchemistry can reflect local stream water chemistry and thus location at the time of accretion (Campana and Thorrold 2001; Wells et al. 2003; Feyrer et al. 2007). While otolith chemistry does reflect ambient chemistry to some extent, other factors can also influence otolith core chemistry and results must be interpreted carefully. For example, physiology, maternal effects, and crystallization can all influence core chemistry (Brophy et al. 2004), especially in species that have yolk sacks like Pacific salmon *Oncorhynchus* spp. (Bacon et al. 2004; Miller and Kent 2009).

More commonly, researchers have assessed fish population structure using genetic markers, such as microsatellites. A smaller number of studies have applied otolith and genetic approaches in parallel. In those studies, the two approaches have provided complementary information about habitat use, movement patterns, and population structure of North American and Australian fishes (Ashford et al. 2006; Bradbury et al. 2008; Barnett-Johnson et al. 2010; Woods et al. 2010; Leis et al. 2011). Chemical tracers can be used to determine whether populations are connected on an ecological time scale, and genetic analyses can be used to determine whether they are connected on an evolutionary time scale. Hence, using multiple types of markers can allow investigators to determine whether populations are connected on multiple time scales (Rubenstein and Hobson 2004; Leis et al. 2011).

The objective of our study was to examine the population structure of a major migratory Neotropical fish, *Prochilodus mariae*, on a regional scale using multiple analytical approaches in parallel. Very little is known about the population structure of migratory fishes in tropical rivers, where many species migrate annually between breeding and feeding areas. South American rivers contain diverse assemblages of migratory fishes, including many long-distance migrants, that support key regional fisheries and are threatened by anthropogenic environmental changes (Barletta et al. 2010). The few existing studies on population structure of migratory fishes in Neotropical rivers suggest wide variation among species, ranging from migratory catfish (Spotted Sorubim *Pseudoplatystoma corruscans*) with high levels of population structure (Pereira et al. 2009) to migratory characins (*Prochilodus* spp.) with no detectable population structure (Hatanaka et al. 2006; Carvalho-Costa et al. 2008). Effective conservation and management of Neotropical fish faunas require a better understanding of their movement patterns and population structure, yet this information is lacking even for most heavily exploited species.

Our focal species, *P. mariae*, and its close relatives are a dominant component of fish biomass and fishery yields in rivers across South America (Barbarino Duque et al. 1998; Sivasundar et al. 2001; Hoeninghaus et al. 2004; Taylor et al. 2006). Prochilodontids typically make extensive annual migrations. *Prochilodus mariae* is iteroparous and migrates yearly from the foothill streams of the eastern slope of the Andes to seasonally inundated floodplains, where they spawn. Field studies have documented the strong influence of *P. mariae* migrations on the structure and functioning of river ecosystems in Venezuela

(Flecker 1996; Taylor et al. 2006; McIntyre et al. 2007), but the implications of these migrations on population structure have not been evaluated. Genetic studies on related species indicate low levels of genetic structure among populations (Turner et al. 2004; Carvalho-Costa et al. 2008). A better understanding of the population biology of *P. mariae* and other fishery species is especially important because these migrants face overfishing (Allan et al. 2005) and increasing degradation of both floodplain and upland habitats caused by deforestation and other human activities, as well as installation of major dams in the Orinoco and Amazon basins (Nilsson et al. 2005).

We coupled otolith chemistry and population genetic methods to address two objectives: (1) infer movement patterns of individual *P. mariae* using otolith microchemical analysis, and determine whether adults using different upland rivers share a common lowland breeding area and (2) determine whether populations of *P. mariae* from different rivers are genetically isolated, or instead disperse and interbreed frequently enough to be managed as a panmictic population.

METHODS

Site Descriptions and Sample Collections

During the 2006 dry season (January–March), we collected *P. mariae* from six river sites in the Apure drainage of the Orinoco basin in Venezuela (Table 1; Figure 1). Some sites were only a few kilometers apart (e.g., Rio Las Marias and Rio Morador), while others were connected to the remainder of the sites only through the Apure River (e.g., Rio Paguey). Two of the sites were on the Rio Portuguesa, which we refer to as Upper Portuguesa, an Andean foothill site, and Lower Portuguesa for a larger downstream site located in the floodplain. All six populations were included in otolith microchemical analysis, but the two downstream sites were excluded from genetic analysis due to sampling constraints (Figure 1; Table 1). Approximately 20 fish were collected from each site by local fishermen using cast nets. Whole fish heads were frozen for later otolith extraction, and muscle samples were preserved in 95% ethanol for DNA analysis.

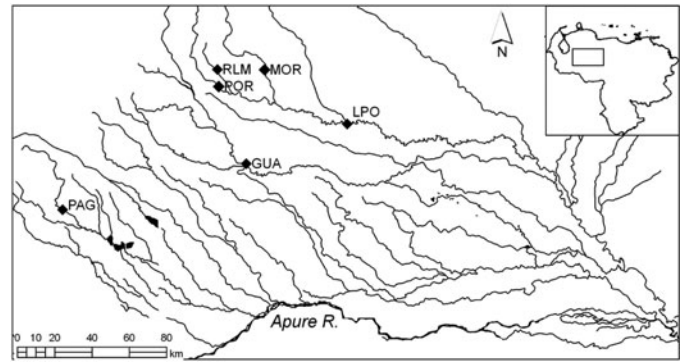


FIGURE 1. Map of the study area showing six collection sites. Sites are abbreviated as follows in all figures and tables: RLM = Rio Las Marias, MOR = Rio Morador, POR = Upper Rio Portuguesa, PAG = Rio Paguey, LPO = Lower Rio Portuguesa Gammareno, GUA = Rio Guanare.

Otolith Microchemical Analysis

We analyzed chemical signatures of the core and edge of the otolith from each individual fish. The edge of the otolith is the most recently accreted layer and reflects the chemistry of the environment inhabited just before collection. We examined otolith edge chemistry to determine whether there is likely to be sufficient chemical variation within the system to separate discrete spawning grounds. Otoliths were analyzed for strontium (Sr), barium (Ba), and magnesium (Mg) and reported as ratios to calcium (Ca). The ratio of stable isotopes of strontium ($^{87}\text{Sr}:$ ^{86}Sr) were also quantified from raw count data.

We extracted the sagittal otolith of each individual using standard techniques (Campana 1999; Bickford and Hannigan 2005) and stored them in plastic tubes until processing. Otoliths were thin-sectioned with a low-speed saw and then polished with fine-grit and velvet polishing pads until the core was distinctly visible. Otoliths were washed with Milli-Q water and sonicated, then analyzed with dry laser ablation (New Wave 213 nm Nd:YAG) dynamic reaction cell-inductively coupled plasma–mass spectrometry (7500c Agilent) at the University of Alaska–Fairbanks. We used a spot size of 20 μm for both edge and core chemistry analyses, which was the smallest spot size with good resolution and was estimated to represent approximately 10 days of otolith accretion. The core region was approximately 45 μm in diameter. Laser energy was set at 100% with

TABLE 1. Study site characteristics. Number of individuals sampled per river, mean SL, and condition indices (SD in parentheses) are described. Otolith and Genetic columns indicate whether those types of analyses were conducted on fish from each site.

| Site | Site type | Number of individuals | SL | Condition index | Otolith | Genetic |
|------------------------|------------|-----------------------|------------|-----------------|---------|---------|
| Rio Guanare (GUA) | Floodplain | 21 | 27.4 (3.5) | 1.48 (0.40) | yes | no |
| Rio Morador (MOR) | Foothill | 16 | 24.8 (1.8) | 1.29 (0.15) | yes | yes |
| Rio Paguey (PAG) | Foothill | 20 | 29.8 (2.1) | 1.27 (0.08) | yes | yes |
| Upper Portuguesa (POR) | Foothill | 20 | | | yes | yes |
| Lower Portuguesa (LPO) | Floodplain | 20 | 25.3 (2.6) | 1.45 (0.14) | yes | no |
| Rio Las Marias (RLM) | Foothill | 25 | 22.0 (2.3) | 1.38 (0.12) | yes | yes |

the frequency at 20 Hz and the number of bursts at 200. Flow of the carrier gas, Ar, was set at 1.05 mL/min. Each sample was ablated using a line width size of 25 μ m and data were averaged.

Results were calibrated against a silicate glass certified reference material (NIST 610) and corrected for gas blanks before every sample. Samples were run in a randomized order and each sample measurement was preceded by a gas blank measurement with recalibration (gas blank and NIST SRM 610) every 10 samples. Background counts were monitored for 20 s before and after each sample. Concentrations of all elements (Ca, Sr, Ba, Mg) were calculated relative to NIST SRM 610 after proper correction for gas blank, matrix, and drift effects. This allowed us to compare elemental abundances relative to Ca content across otolith samples (Campana 1999). All measurements were greater than the limit of detection and external precision (i.e., relative standard deviations) was better than 3% for all elements and for each isotope of Sr.

Microsatellite DNA Genotyping

We developed a microsatellite DNA library for *P. mariae* (Barker et al. 2011) and genotyped approximately 20 individuals from each site using 11 microsatellite loci (Table 2). We used a universal tag method to generate fluorescently labeled

PCR products for genotyping (Schuelke 2000). We analyzed labeled PCR products on a 3100 Genetic Analyzer (Applied Biosystems) and determined allele sizes using GENEMAPPER version 3.0 (Applied Biosystems).

Statistical Analyses

Otolith microchemistry.—We used quadratic discriminant function analysis (QDFA) and a jackknife procedure to cross validate chemical signatures found at each site based on edge and core data. First, we used QDFA to classify fish into groups based on edge chemistry. This classification was expected to match the known collection sites. Since many freshwater systems are geochemically similar, edge chemical classification allowed us to evaluate whether there was sufficient background variation in chemistry among rivers to distinguish sites. We used multivariate analysis of variance (MANOVA) to determine whether multivariate chemical signatures were significantly different among sites and analysis of variance (ANOVA) to identify differences for individual elemental signatures differed among sites for both core and edge chemistry.

We performed an identical QDFA analysis using the core chemical data to determine whether individuals from each site shared a chemically similar natal environment. Since *P. mariae*

TABLE 2. Characteristics of 11 *P. mariae* microsatellite loci, including locus name, GenBank accession number, repeat motif, fluorescent dye color, primer sequence (F: forward primer, R: reverse primer), annealing temperature (T_a), $MgCl_2$ concentration, allele size range (bp: base pairs), number of individuals genotyped (n), number of alleles (N_A), number of private alleles, observed heterozygosity (H_o) and expected heterozygosity (H_e).

| Locus and GenBank accession Number | Repeat motif and fluorescent dye color | Primer sequence (5'-3') | T_a (°C) | $MgCl_2$ (mM) | Allele size range (bp) | Number | | | | |
|------------------------------------|--|--|------------|---------------|------------------------|--------|-------|-----------------|-------|-------|
| | | | | | | n | N_A | private alleles | H_o | H_e |
| Pmar01 JF832396 | (AGAT) ₂₁ VIC | F: CCACAAGAAGATAGACCCAACGAT R: TTTCATATACCGGTCTGCCTTTCT | 62 | 1.6 | 280–413 | 74 | 33 | 10 | 0.95 | 0.96 |
| Pmar06 JF832398 | (TATC) ₁₆ FAM | F: CATGTCCTTGAATTTCCCTCTGTC R: CTGGTTGATCTGGCTCTTGCTGGTA | 62 | 1.6 | 118–269 | 72 | 34 | 15 | 0.58 | 0.96 |
| Pmar07 JF832400 | (AGAT) ₁₄ FAM | F: GCCCTCCCTTAGTCACCCTGTTA R: AATACCAAATTTATAAGTTGCGTCCACA | 62 | 2.0 | 321–405 | 73 | 24 | 10 | 0.52 | 0.93 |
| Pmar11 JF832395 | (GTT) ₁₁ PET | F: ATAATAAATCCAGGTGCAGCTTCAGTCA R: CTTCAGAGTCCACATCACCATCAGTAAC | 62 | 2.0 | 190–226 | 75 | 11 | 1 | 0.87 | 0.84 |
| Pmar17 JF832405 | (TATC) ₂₀ NED | F: TAGGCAATGTGACTGAGTGTGAGT R: GAGCCGTTTAATAACAGCCTACAG | 62 | 1.6 | 196–302 | 75 | 19 | 7 | 0.93 | 0.91 |
| Pmar19 JF832404 | (TTG) ₁₀ FAM | F: CAGGCATGTACCAACCCAAACC R: ATGAGTCAAATAACGGCACAGG | 62 | 1.6 | 270–325 | 75 | 17 | 6 | 0.76 | 0.88 |
| Pmar20 JF832402 | (CAA) ₁₃ NED | F: TTCCAGCACCAGATAAACCA R: GGCTAAACCTGCTCGTCCTCT | 62 | 1.6 | 191–237 | 75 | 17 | 3 | 0.68 | 0.90 |
| Pmar22 JF832397 | (TTG) ₁₅ VIC | F: AGCCTGACGCTAGTCCTTTTCTT R: GTATAATCCGGCTGTTTTAGTTCAATA | 62 | 1.6 | 187–238 | 75 | 33 | 8 | 0.96 | 0.95 |
| Pmar27 JF832399 | (AAC) ₁₈ FAM | F: GGTCAAAAATCCCATAAACACAT R: TATAGTAAACAAGAACAATCAAAGAGA | 60 | 1.6 | 292–378 | 73 | 26 | 9 | 0.58 | 0.89 |
| Pmar28 JF832401 | (GTT) ₁₉ NED | F: ATATTTGTTTTGTCCATTCATCA R: CACTGGGTATTTACTTTGTTATGTG | 60 | 1.6 | 198–239 | 68 | 15 | 2 | 0.39 | 0.90 |
| Pmar29 JF832403 | (TC) ₁₈ VIC | F: ACTGAGTGGCGACTGGTCCTG R: CTCTGAGTCAGCTGACGGTGCTC | 60 | 1.6 | 143–172 | 75 | 14 | 3 | 0.87 | 0.87 |

reproduce while migrating downstream (Barbarino Duque et al. 1998), it is unlikely that the natal environments are the same as the collection environments. Hence, we were only able to explore whether natal sites were distinct and shared among individuals from a population. We conducted all otolith chemistry statistical analyses using R (R Development Core Team 2010).

Microsatellite DNA.—Tests for Hardy–Weinberg equilibrium and linkage disequilibrium were performed using GENEPOP version 3.4 (Raymond and Rousset 1995). We used F_{ST} as a measure of population differentiation and calculated pairwise F_{ST} estimates using Arlequin version 3.1 (Excoffier et al. 2005). Genetic structure was also analyzed using Bayesian clustering methods in two programs to ensure the consistency and hence reliability of the results: STRUCTURE version 2.2 (Pritchard et al. 2000) and GENELAND version 2.0.12 (Guillot et al. 2005). Both STRUCTURE and GENELAND use a Markov Chain Monte Carlo (MCMC) approach to infer the most likely number of genetic groups (K). The cluster analysis in GENELAND uses GPS coordinates of collection sites to account for isolation by distance, which can increase detection of low levels of population structure.

In STRUCTURE, we performed MCMC runs with values of K set to 1 to 10 to determine the most likely number of genetic groups. We chose that range for K because we wanted to include all likely numbers of genetic populations for our four geographic populations. For each value of K , we conducted 10 independent runs using the admixture model, the correlated allele frequency model, and alphapropsd set to 0.001 (higher values led to substantial variations of alpha along the runs).

For each run, a burn-in period of 50,000 steps followed by one million steps of data collection was sufficient to ensure MCMC convergence. To infer the most likely number of genetic groups in the data set, we looked for the value of K that gave the highest $\ln P(D)$ (the most appropriate method with the type of plot of $\ln P(D)$ as a function of K we obtained [see Results, Pritchard et al. 2000; Evanno et al. 2005]).

In GENELAND, we performed 10 runs of 1,000,000 MCMC iterations of the spatial model, where K was allowed to vary between 1 and 10 (number initialized to 3) with the following set of parameters that were chosen after conducting preliminary runs to adjust their values: maximum rate of the Poisson process used to generate the Voronoi cells = 500, maximum number of nuclei in the Poisson–Voronoi tessellation = 500, allele frequency model = Dirichlet. The uncertainty associated with the spatial coordinates of specimen collection was 100 m. The most likely K was inferred as the modal number of genetic groups estimated among 1,000,000 iterations of the best of the 10 runs. To select the best run, we used the posterior density of the runs (after a burn-in of 200,000 iterations) as an estimator of their quality: the posterior density is estimated for each state of the parameters, explored along the Markov chain, and represents the posterior probability of that current state of parameters.

RESULTS

Otolith Microchemistry

Cross validation of otolith edge chemistry resulted in moderate to good classification of individuals to each collection site; 65–80% of individuals were correctly classified (Table 3). The

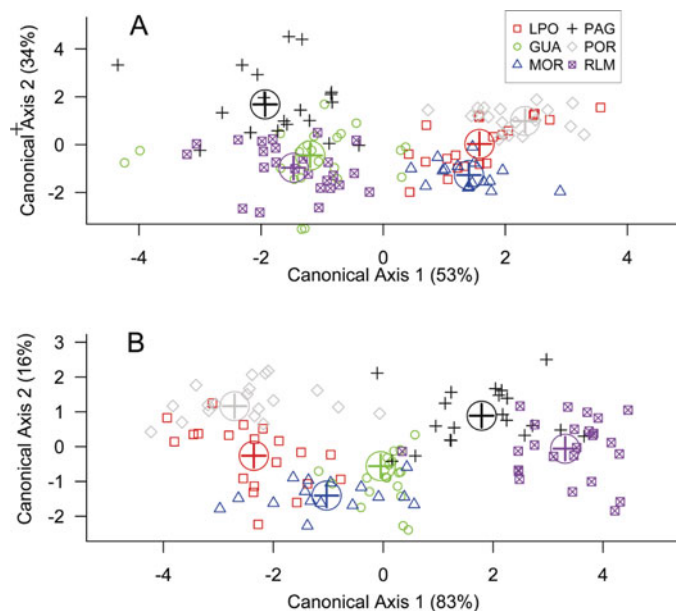


FIGURE 2. Plots demonstrating spatial variation in multivariate otolith chemistry (canonical axes are a compilation of Sr:Ca, Ba:Ca, Mg:Ca, and $^{87}\text{Sr}:$ ^{86}Sr chemical signatures). Panel (A) shows variation in otolith core chemistry and Panel (B) shows variation in otolith edge chemistry. Each small point represents an individual fish and the large points represent the centroid for each site group. [Figure available online in color.]

TABLE 3. Classification results from QDFA for otolith edges (top) and cores (bottom). Table shows the percentage of individuals from each collection site that are classified into each site. Total numbers of individuals collected are shown in parentheses next to the collection site label. Numbers in bold italics indicate that the collection and classification sites are the same.

| Collection site | Classification site | | | | | |
|-----------------|---------------------|-------------|-------------|-------------|-------------|-------------|
| | PAG (%) | RLM (%) | GUA (%) | MOR (%) | LPO (%) | POR (%) |
| | Edge | | | | | |
| PAG (20) | 65.0 | 25.0 | 5.0 | 0.0 | 0.0 | 5.0 |
| RLM (26) | 16.0 | 80.0 | 4.0 | 0.0 | 0.0 | 0.0 |
| GUA (21) | 4.8 | 0.0 | 76.2 | 14.3 | 4.8 | 0.0 |
| MOR (16) | 0.0 | 0.0 | 25.0 | 68.8 | 6.3 | 0.0 |
| LPO (20) | 0.0 | 0.0 | 5.0 | 20.0 | 60.0 | 15.0 |
| POR (20) | 0.0 | 0.0 | 10.0 | 0.0 | 15.0 | 75.0 |
| | Core | | | | | |
| PAG (20) | 70.0 | 5.0 | 25.0 | 0.0 | 0.0 | 0.0 |
| RLM (26) | 0.0 | 80.0 | 20.0 | 0.0 | 0.0 | 0.0 |
| GUA (21) | 19.0 | 33.3 | 38.1 | 4.8 | 4.8 | 0.0 |
| MOR (16) | 0.0 | 0.0 | 0.0 | 81.3 | 18.8 | 0.0 |
| LPO (20) | 0.0 | 0.0 | 5.0 | 35.0 | 20.0 | 40.0 |
| POR (20) | 0.0 | 0.0 | 0.0 | 0.0 | 25.0 | 75.0 |

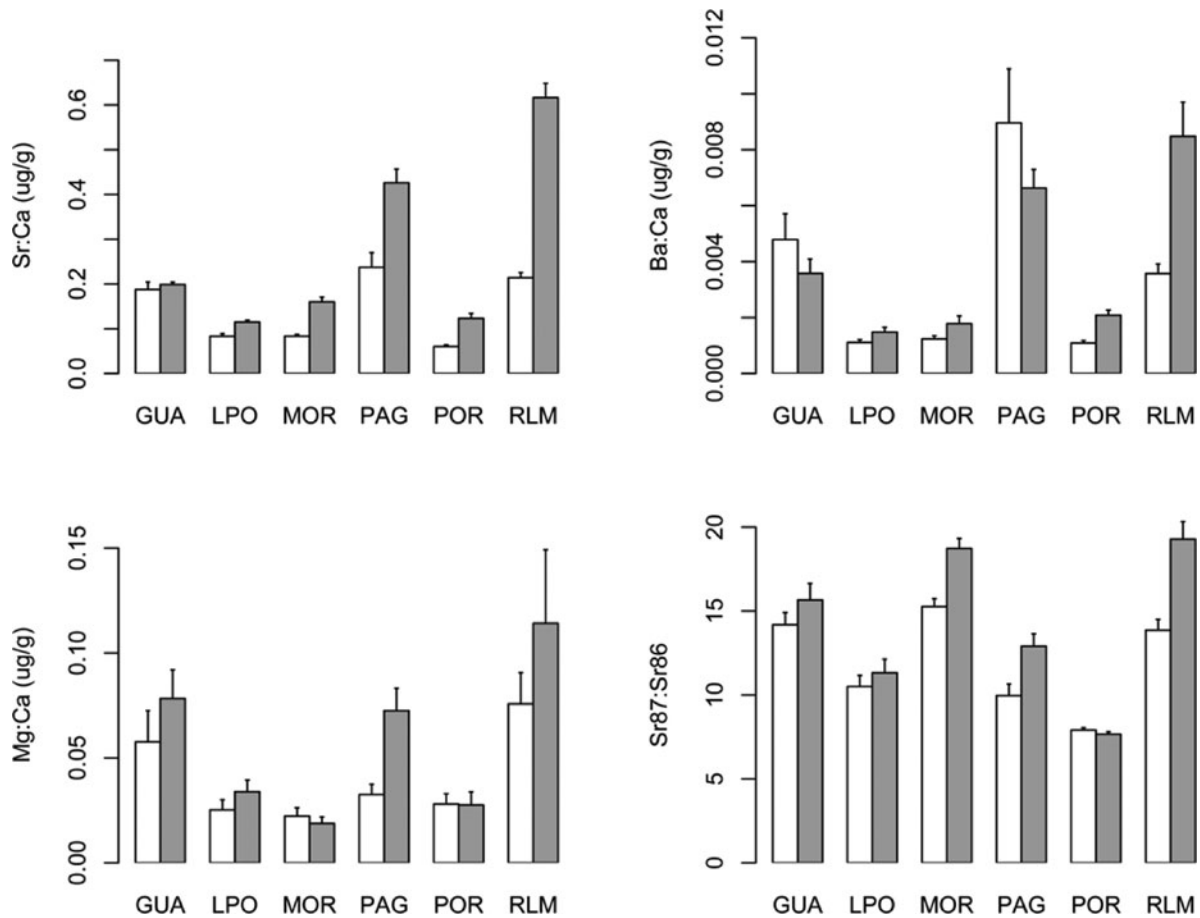


FIGURE 3. Elemental ratios from the core and edge of otoliths collected at each river. Site codes are specified in Figure 1. Error bars show the confidence interval. White bars represent core chemistry and grey bars represent edge chemistry.

MANOVA results indicate that otolith chemical signatures differed significantly among the six sites for both edge and core chemistry (Figure 2; Core Chemistry: Pillai's trace statistic = 1.6093, $F = 15.617$, $df = 20$, $P < 0.001$; Edge Chemistry: Pillai's trace statistic = 1.4683, $F = 13.456$, $df = 20$, $P < 0.001$). All individual element signatures were also significantly different among sites (Figure 3; Table 4). The most frequent misclassifications were of fish collected from two sites along the same river (Upper Portuguesa and Lower Portuguesa), where edge chemical signatures were similar. Based on chemical variation among otolith cores, a high proportion of individuals from each of the smaller, foothill rivers (Paguey, Las Marias, Upper Portuguesa, Morador) were classified by cross validation as sharing a natal site distinct from those of other populations. In contrast, fish from the larger, downstream rivers (Guanare, Lower Portuguesa) showed wide variation in natal site classifications (Figure 2; Table 3).

Microsatellite DNA

Our microsatellite markers showed high levels of genetic diversity (11–34 alleles per locus). Genotype data showed no significant evidence of gametic disequilibrium or deviation from

Hardy–Weinberg proportions. In addition, there was no significant difference between observed and expected heterozygosity for microsatellite loci. Pairwise F_{st} values indicate extremely low levels of population differentiation (Table 5).

Both Bayesian clustering analyses indicate that all populations comprise a single genetic group; there was no detectable population structure at the spatial scale of our study

TABLE 4. Results of ANOVA indicate differences among sites for each elemental signature in the otolith. Otolith core and edge results for the four elemental signatures are presented.

| Elemental ratio | Otolith region | F -value | P -value |
|-----------------|----------------|------------|------------|
| Sr:Ca | Edge | 104.36 | <0.001 |
| Sr:Ca | Core | 60.837 | <0.001 |
| Mg:Ca | Edge | 13.668 | <0.001 |
| Mg:Ca | Core | 7.085 | <0.001 |
| Ba:Ca | Edge | 45.583 | <0.001 |
| Ba:Ca | Core | 51.448 | <0.001 |
| Sr 87:86 | Edge | 26.179 | <0.001 |
| Sr 87:86 | Core | 24.44 | <0.001 |

TABLE 5. Pairwise F_{ST} values for the four sites used in genetic analyses.

| Site | Site | | | |
|------|---------|---------|---------|---------|
| | RLM | PAG | POR | MOR |
| RLM | 0.00000 | | | |
| PAG | 0.00326 | 0.00000 | | |
| POR | 0.00822 | 0.00946 | 0.00000 | |
| MOR | 0.00246 | 0.00696 | 0.00856 | 0.00000 |

(Figure 4). Analysis in STRUCTURE showed that all individuals are most likely drawn from a single genetic population ($K = 1$). GENELAND also indicated a single population ($K = 1$). These results indicated that gene flow in this species was sufficient to prevent genetic differentiation at the spatial scale of our collections.

DISCUSSION

Otolith chemical analyses suggest that most individuals from each collection site share a distinct, common natal site, but mi-

cro-satellite genetic analyses did not detect any genetic differentiation or structure among populations of *P. mariae*. Based on otolith edge chemical data, we were able to correctly classify 68–80% of individuals to the river where they were collected, which is relatively high classification success compared to similar studies in other freshwater ecosystems (Feyrer et al. 2007). Differences in otolith edge chemistries among sites implied environmental heterogeneity within the system and gave confidence that otolith core chemistry might differ should different river populations exploit geographically different spawning sites. Our results show that a high proportion of individuals (70–81%) from the four upstream collection sites exhibit distinctive natal signatures (Table 3), implying that breeding areas for these stocks are largely distinct.

Despite the otolith chemistry evidence for distinct breeding areas, there was sufficient genetic mixture among populations to prevent population genetic differentiation. The simplest explanation for this pattern is that the level of mixing among populations overrides the genetic structure arising from spatial segregation of breeding. Our assignments based on QDFA indicate up to 30% mixing of individuals, i.e., up to 30% individuals were assigned to a different natal site than the site in which they were collected. If this level of mixing carries through to the breeding season, then it would be no surprise that *P. mariae* are effectively panmictic within the study region. In addition, since this species breeds following the downstream migration, interbreeding among populations may easily occur during passage downriver. *Prochilodus mariae* are iteroparous migrants, so individual fish also have multiple opportunities to spawn during the course of their lifetime, including the possibility that fish that return to a river that has different chemistry than their natal site may continue to associate with a nonnatal population and contribute to genetic mixing over many breeding seasons.

The data we collected for this study do not allow us to differentiate between these mechanisms for genetic homogenization, but future studies of *P. mariae* or other migratory fishes could collect data along a full ontogenetic transect of otoliths to investigate whether individual fish migrate to breed and feed in the same locations throughout their lives (Limburg et al. 2011). Additional sampling during the breeding season may also provide key information about whether population structure exists. If individuals return to breed in the same location every year, populations may have stronger genetic structure if they were sampled during the breeding season and a comparison of population structure before and after breeding could be very informative.

Both the spatial patterns inferred from otolith microchemistry and the lack of genetic structure should be important considerations for the management of the heavily exploited *P. mariae* fishery. Because at least some interbreeding occurs among populations, migrants from nearby populations could help to maintain genetic diversity and breeding populations in the face of heavy exploitation. However, the dominance of spatial segregation of breeding sites evident from otolith chemistry indicates that while there is enough mixing to prevent genetic

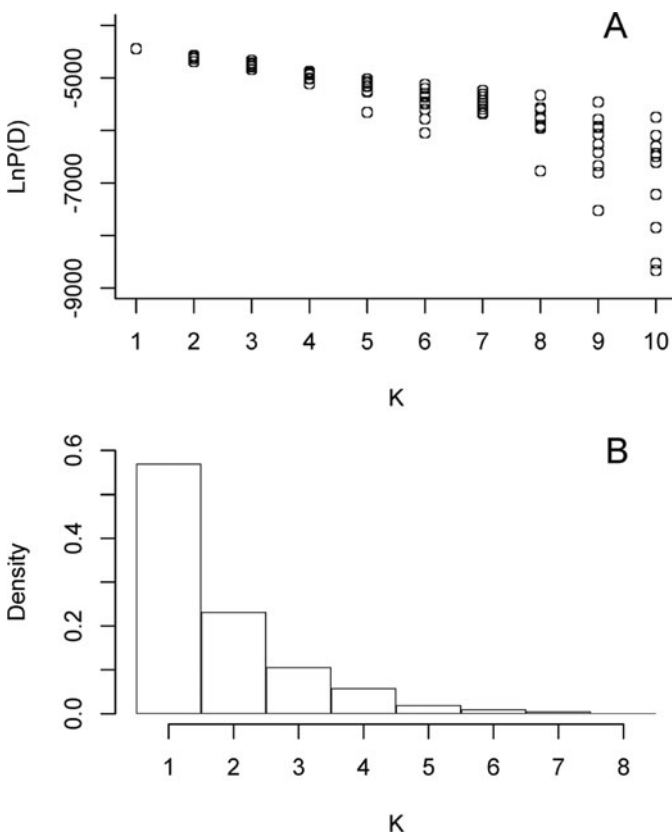


FIGURE 4. Results from genetic analysis in STRUCTURE and GENELAND. Panel (A) shows STRUCTURE results. Values on the y-axis are the estimated log of the probability of the data and show a relative estimate of the fit of the clustering solutions to the data for each number of genetic groups (K). Panel (B) shows results from genetic analysis in GENELAND. Each bar represents the proportion of MCMC iterations that inferred that number of genetic groups (K).

differentiation, it is probably too little to support a heavily harvested population by immigration alone. Thus, severe overfishing or habitat destruction in either breeding or feeding areas used by each stock is likely to strongly diminish its viability.

Our results suggest that tracer methods, such as otolith microchemical analyses, are especially valuable in systems where there is little or no population structure. For *P. mariae*, the molecular markers we used did not allow us to detect low levels of population structure. The limited geographic scale of our study did not allow a complete assessment of genetic structure for the species, but other studies also found low levels of population structure in prochilodontids. For instance, a case study of *Prochilodus costatus* (Carvalho-Costa et al. 2008) and a broad review of migratory fish biogeography in South America (Turner et al. 2004) both indicated that individual prochilodontid movements are very widespread and populations are genetically homogenous. In addition, our sampling included one river (Rio Paguey) that is disconnected from the remainder of the sites by the Rio Apure (a large tributary to the Orinoco River). Nevertheless, fish from the Rio Paguey were part of the same genetic group as the other sites. Otolith chemical data allowed us to detect population structure on the spatial scale of our study sites even though no genetic differences existed.

Relatively few studies have used both otolith and genetic markers in concert to study migratory fish population biology, but some have gleaned additional insight from using both methods (Miller et al. 2005; Baerwald et al. 2007; Feyrer et al. 2007; Bradbury et al. 2008; Barnett-Johnson et al. 2010; Woods et al. 2010). In other systems, coupled otolith and genetic approaches have mostly produced complementary results; specifically, genetic and otolith approaches both indicate that population boundaries occur in the same location, with few exceptions (Thorrold et al. 2001).

Our analysis of *P. mariae* suggests that different markers may indicate different levels of connectivity: we did not detect genetic structure despite identifying geographic structure using otolith chemical markers. The lack of genetic structure in our system does not necessarily contradict otolith chemical results but does indicate that it may be necessary to consider both ecological connectivity (e.g., otolith chemistry) and evolutionary connectivity (e.g., microsatellites) to understand systems with extremely low levels of population structure. Other coupled methodological combinations besides otolith and genetic markers also show promise for understanding animal movements; previous studies have used genetic and physiological markers to provide insight into fish migrations (Cooke et al. 2008), and others have coupled stable isotope markers with genetics to understand bird and butterfly migrations (Rubenstein and Hobson 2004).

Our results provide basic information on the population biology of one important migratory Neotropical species, but additional study is needed to effectively manage and conserve *P. mariae* and other migratory Neotropical fishes in the face of habitat destruction and overharvest (Barbarino Duque et al.

1998; Mateus et al. 2004). Population structure has been strong and easy to detect for some South American species (Pereira et al. 2009), but most genetic analyses have identified high levels of genetic diversity and very low levels of genetic structure (Hatanaka et al. 2006; Carvalho-Costa et al. 2008; Calcagnotto and DeSalle 2009; Iervolino et al. 2010). Despite the ubiquity and utility of variable genetic markers to analyze population structure (Hudson 2008), our results suggest that genetic data should not be treated as definitive when no population structure is detected. Coupling direct methods of tracking individual animals (e.g., telemetry) with indirect methods of assessing population connectivity (e.g., genetics) is likely to produce robust information about population structure that will be appropriate for the limited temporal and spatial scales of many management efforts. In the case of *P. mariae*, populations are managed by local communities on a small scale, so understanding whether populations are connected may have important implications for whether local stewardship and resource management will actually result in a more sustainable fishery. Our case study of *P. mariae* demonstrates the complex nature of understanding population structure.

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REFERENCES

- Allan, J. D., R. Abell, Z. Hogan, C. Revenga, B. W. Taylor, R. L. Welcomme, and K. Winemiller. 2005. Overfishing of inland waters. *BioScience* 55:1041–1051.
- Alvarado Bremer, J. R., J. Mejuto, J. Gómez-Márquez, F. Boán, P. Carpintero, J. M. Rodríguez, J. Viñas, T. W. Greig, and B. Ely. 2005. Hierarchical analyses of genetic variation of samples from breeding and feeding grounds confirm the genetic partitioning of northwest Atlantic and South Atlantic populations of Swordfish (*Xiphias gladius* L.). *Journal of Experimental Marine Biology and Ecology* 327:167–182.
- Ashford, J. R., A. I. Arkhipkin, and C. M. Jones. 2006. Can the chemistry of otolith nuclei determine population structure of Patagonian Toothfish *Dissostichus eleginoides*? *Journal of Fish Biology* 69:708–721.
- Bacon, C. R., P. K. Weber, K. A. Larsen, R. Reisenbichler, J. A. Fitzpatrick, and J. L. Wooden. 2004. Migration and rearing histories of Chinook Salmon (*Oncorhynchus tshawytscha*) determined by ion microprobe Sr isotope and Sr/Ca transects of otoliths. *Canadian Journal of Fisheries and Aquatic Sciences* 61:2425–2439.
- Baerwald, M., V. Bien, F. Feyrer, and B. May. 2007. Genetic analysis reveals two distinct Sacramento splittail (*Pogonichthys macrolepidotus*) populations. *Conservation Genetics* 8:159–167.
- Barbarino Duque, A., D. C. Taphorn, and K. O. Winemiller. 1998. Ecology of the Coporo, *Prochilodus mariae* (Characiformes, Prochilodontidae), and

- status of annual migrations in western Venezuela. *Environmental Biology of Fishes* 53:33–46.
- Barker, F. K., J. J. Bell, S. M. Bogdanowicz, S. L. Bonatto, F. Cezilly, S. M. Collins, C. Dubreuil, M. J. Dufort, C. Eraud, R. Fuseya, E. A. Heap, N. Jacobsen, M. Madders, R. McEwing, A. P. Michel, F. Mougeot, R. S. Ogden, L. C. Orantes, A. S. Othman, E. Parent, P. Pulido-Santacruz, R. Rioux-Paré, M. F. Roberts, R. Rosazlina, T. Sakamoto, P. S. DeLeón, J. M. Sévigny, P. St. Onge, J. Terraube, R. E. Tingay, R. Tremblay, S. Watanabe, and R. A. Wattier. 2011. Permanent genetic resources added to molecular ecology resources database 1 June 2011–31 July 2011. *Molecular Ecology Resources* 11:1124–1126.
- Barletta, M., A. J. Jaureguizar, C. Baigun, N. F. Fontoura, A. A. Agostinho, V. M. F. Almeida-Val, A. L. Val, R. A. Torres, L. F. Jimenes-Segura, T. Giarrizzo, N. N. Fabrè, V. S. Batista, C. Lasso, D. C. Taphorn, M. F. Costa, P. T. Chaves, J. P. Vieira, and M. F. M. Corrêa. 2010. Fish and aquatic habitat conservation in South America: a continental overview with emphasis on neotropical systems. *Journal of Fish Biology* 76:2118–2176.
- Barnett-Johnson, R., D. J. Teel, and E. Casillas. 2010. Genetic and otolith isotopic markers identify salmon populations in the Columbia River at broad and fine geographic scales. *Environmental Biology of Fishes* 89:533–546.
- Bickford, N., and R. Hannigan. 2005. Stock identification of Walleye via otolith chemistry in the Eleven Point River, Arkansas. *North American Journal of Fisheries Management* 25:1542–1549.
- Bradbury, I. R., S. E. Campana, and P. Bentzen. 2008. Estimating contemporary early life-history dispersal in an estuarine fish: integrating molecular and otolith elemental approaches. *Molecular Ecology* 17:1438–1450.
- Brophy, D., T. E. Jeffries, and B. S. Danilowicz. 2004. Elevated manganese concentrations at the cores of clupeid otoliths: possible environmental, physiological, or structural origins. *Marine Biology* 144:779–786.
- Calcagnotto, D., and R. DeSalle. 2009. Population genetic structuring in pacu (*Piaractus mesopotamicus*) across the Paraná-Paraguay basin: evidence from microsatellites. *Neotropical Ichthyology* 7:607–616.
- Campana, S. E. 1999. Chemistry and composition of fish otoliths: pathways, mechanisms and applications. *Marine Ecology Progress Series* 188:263–297.
- Campana, S. E., and S. R. Thorrold. 2001. Otoliths, increments, and elements: keys to a comprehensive understanding of fish populations? *Canadian Journal of Fisheries and Aquatic Sciences* 58:30–38.
- Carvalho-Costa, L. F., T. Hatanaka, and P. M. Galetti Jr. 2008. Evidence of lack of population substructuring in the Brazilian freshwater fish *Prochilodus costatus*. *Genetics and Molecular Biology* 31:377–380.
- Cooke, S. J., S. G. Hinch, A. P. Farrell, D. A. Patterson, K. Miller-Saunders, D. W. Welch, M. R. Donaldson, K. C. Hanson, G. T. Crossin, M. T. Mathes, A. G. Lotto, K. A. Hruska, I. C. Olsson, G. N. Wagner, R. Thomson, R. Hourston, K. K. English, S. Larsson, J. M. Shrimpton, and G. Van der Kraak. 2008. Developing a mechanistic understanding of fish migrations by linking telemetry with physiology, behavior, genomics and experimental biology: an interdisciplinary case study on adult Fraser River Sockeye Salmon. *Fisheries* 33:321–339.
- Evanno, G., S. Regnaut, and J. Goudet. 2005. Detecting the number of clusters of individuals using the software STRUCTURE: a simulation study. *Molecular Ecology* 14:2611–2620.
- Excoffier, L., G. Laval, and S. Schneider. 2005. Arlequin (version 3.0): an integrated software package for population genetics data analysis. *Evolutionary Bioinformatics Online* 1:47–50.
- Feyrer, F., J. Hobbs, M. Baerwald, T. Sommer, Q. Z. Yin, K. Clark, B. May, and W. Bennett. 2007. Otolith microchemistry provides information complementary to microsatellite DNA for a migratory fish. *Transactions of the American Fisheries Society* 136:469–476.
- Flecker, A. S. 1996. Ecosystem engineering by a dominant detritivore in a diverse tropical stream. *Ecology* 77:1845–1854.
- Guillot, G., F. Mortier, and A. Estoup. 2005. GENELAND: a computer package for landscape genetics. *Molecular Ecology Notes* 5:712–715.
- Hatanaka, T., F. Henrique-Silva, and P. M. Galetti Jr. 2006. Population substructuring in a migratory freshwater fish *Prochilodus argenteus* (Characiformes, Prochilodontidae) from the São Francisco River. *Genetica* 126:153–159.
- Heath, M. R., P. A. Kunzlik, A. Gallego, S. J. Holmes, and P. J. Wright. 2008. A model of meta-population dynamics for North Sea and west of Scotland cod: the dynamic consequences of natal fidelity. *Fisheries Research* 93:92–116.
- Hilborn, R., T. P. Quinn, D. E. Schindler, and D. E. Rogers. 2003. Biocomplexity and fisheries sustainability. *Proceedings of the National Academy of Sciences of the USA* 100:6564–6568.
- Hoinghaus, D. J., K. O. Winemiller, and D. C. Taphorn. 2004. Compositional change in fish assemblages along the Andean piedmont–Llanos floodplain gradient of the Río Portuguesa, Venezuela. *Neotropical Ichthyology* 2:85–92.
- Hudson, M. E. 2008. Sequencing breakthroughs for genomic ecology and evolutionary biology. *Molecular Ecology Resources* 8:3–17.
- Iervolino, F., E. Kawakami de Resende, and A. W. S. Hilsdorf. 2010. The lack of genetic differentiation of pacu (*Piaractus mesopotamicus*) populations in the upper-Paraguay basin revealed by the mitochondrial DNA D-loop region: implications for fishery management. *Fisheries Research* 101:27–31.
- Kell, L. T., M. Dickey-Collas, N. T. Hintzen, R. D. M. Nash, G. M. Pilling, and B. A. Roel. 2009. Lumpers or splitters? evaluating recovery and management plans for metapopulations of herring. *ICES Journal of Marine Science* 66:1776–1783.
- Leis, J. M., L. Van Herwerden, and H. M. Patterson. 2011. Estimating connectivity in marine fish populations: what works best? *Oceanography and Marine Biology: An Annual Review* 49:193–234.
- Limburg, K. E., C. Olson, Y. Walther, D. Dale, C. P. Slomp, and H. Høie. 2011. Tracking Baltic hypoxia and cod migration over millennia with natural tags. *Proceedings of the National Academy of Sciences of the USA* 108:E177–E182.
- Mateus, L. A. F., J. M. F. Penha, and M. Petrere. 2004. Fishing resources in the Rio Cuiabá basin, Pantanal do Mato Grosso, Brazil. *Neotropical Ichthyology* 2:217–227.
- Mather, M. E., J. T. Finn, S. M. Pautzke, D. Fox, T. Savoy, H. M. Brundage III, L. A. Deegan, and R. M. Muth. 2010. Diversity in destinations, routes and timing of small adult and sub-adult Striped Bass *Morone saxatilis* on their southward autumn migration. *Journal of Fish Biology* 77:2326–2337.
- McDowall, R. M. 2001. Anadromy and homing: two life-history traits with adaptive synergies in salmonid fishes? *Fish and Fisheries* 2:78–85.
- McIntyre, P. B., L. E. Jones, A. S. Flecker, and M. J. Vanni. 2007. Fish extinctions alter nutrient recycling in tropical freshwaters. *Proceedings of the National Academy of Sciences of the USA* 104:4461–4466.
- Miller, J. A., M. A. Banks, D. Gomez-Uchida, and A. L. Shanks. 2005. A comparison of population structure in Black Rockfish (*Sebastes melanops*) as determined with otolith microchemistry and microsatellite DNA. *Canadian Journal of Fisheries and Aquatic Sciences* 62:2189–2198.
- Miller, J. A., and A. J. R. Kent. 2009. The determination of maternal run time in juvenile Chinook Salmon (*Oncorhynchus tshawytscha*) based on Sr/Ca and ⁸⁷Sr/⁸⁶Sr within otolith cores. *Fisheries Research* 95:373–378.
- Nilsson, C., C. A. Reidy, M. Dynesius, and C. Revenga. 2005. Fragmentation and flow regulation of the world's large river systems. *Science* 308:405–408.
- Palkovacs, E. P., K. B. Dion, D. M. Post, and A. Caccone. 2008. Independent evolutionary origins of landlocked Alewife populations and rapid parallel evolution of phenotypic traits. *Molecular Ecology* 17:582–597.
- Pereira, L. H. G., F. Foresti, and C. Oliveira. 2009. Genetic structure of the migratory catfish *Pseudoplatystoma corruscans* (Siluriformes: Pimelodidae) suggests homing behaviour. *Ecology of Freshwater Fish* 18:215–225.
- Pritchard, J. K., M. Stephens, and P. Donnelly. 2000. Inference of population structure using multilocus genotype data. *Genetics* 155:945–959.
- R Development Core Team. 2010. R: a language and environment for statistical computing. R Foundation for Statistical Computing, Vienna.
- Raymond, M., and F. Rousset. 1995. GENEPOP (version 1.2): population genetics software for exact tests and ecumenicism. *Journal of Heredity* 86:248–249.

- Rubenstein, D. R., and K. A. Hobson. 2004. From birds to butterflies: animal movement patterns and stable isotopes. *Trends in Ecology and Evolution* 19:256–263.
- Ruzzante, D. E., S. Mariani, D. Bekkevold, C. André, H. Mosegaard, L. A. W. Clausen, T. G. Dahlgren, W. F. Hutchinson, E. M. C. Hatfield, E. Torstensen, J. Brigham, E. J. Simmonds, L. Laikre, L. C. Larsson, R. J. M. Stet, N. Ryman, and G. R. Carvalho. 2006. Biocomplexity in a highly migratory pelagic marine fish, Atlantic Herring. *Proceedings of the Royal Society of London B* 273:1459–1464.
- Schindler, D. E., R. Hilborn, B. Chasco, C. P. Boatright, T. P. Quinn, L. A. Rogers, and M. S. Webster. 2010. Population diversity and the portfolio effect in an exploited species. *Nature* 465:609–612.
- Schuelke, M. 2000. An economic method for the fluorescent labeling of PCR fragments. *Nature Biotechnology* 18:233–234.
- Shaklee, J. B., and P. Bentzen. 1998. Genetic identification of stocks of marine fish and shellfish. *Bulletin of Marine Science* 62:589–621.
- Sivasundar, A., E. Bermingham, and G. Ortí. 2001. Population structure and biogeography of migratory freshwater fishes (*Prochilodus*: Characiformes) in major South American rivers. *Molecular Ecology* 10: 407–417.
- Taylor, B. W., A. S. Flecker, and R. O. Hall Jr. 2006. Loss of a harvested fish species disrupts carbon flow in a diverse tropical river. *Science* 313:833–836.
- Thorold, S. R., C. Latkoczy, P. K. Swart, and C. M. Jones. 2001. Natal homing in a marine fish metapopulation. *Science* 291:297–299.
- Turner, T. F., M. V. McPhee, P. Campbell, and K. O. Winemiller. 2004. Phylogeography and intraspecific genetic variation of prochilodontid fishes endemic to rivers of northern South America. *Journal of Fish Biology* 64:186–201.
- Waldman, J., C. Grunwald, and I. Wirgin. 2008. Sea Lamprey *Petromyzon marinus*: an exception to the rule of homing in anadromous fishes. *Biology Letters* 4:659–662.
- Wells, B. K., B. E. Rieman, J. L. Clayton, D. L. Horan, and C. M. Jones. 2003. Relationships between water, otolith, and scale chemistries of Westslope Cutthroat Trout from the Coeur d'Alene River, Idaho: the potential application of hard-part chemistry to describe movements in freshwater. *Transactions of the American Fisheries Society* 132:409–424.
- Woods, R. J., J. I. Macdonald, D. A. Crook, D. J. Schmidt, and J. M. Hughes. 2010. Contemporary and historical patterns of connectivity among populations of an inland river fish species inferred from genetics and otolith chemistry. *Canadian Journal of Fisheries and Aquatic Sciences* 67:1098–1115.